AN INTEGRATED APPROACH TO MANAGING AND MITIGATING THE RISK OF AGRICULTURAL NONPOINT SOURCE PESTICIDE POLLUTION TO THE AQUATIC ENVIRONMENT

J.M. Dabrowski

Volume 2: Development of risk maps and a risk indicator for identifying hotspots and prioritising risks of pesticide use to aquatic ecosystem health

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EXECUTIVE SUMMARY

BACKGROUND

Agricultural nonpoint source pollution (NPS) plays a major role in the degradation of water quality, particularly regarding pesticides, nutrients and sediment. It has therefore become increasingly important to address the contributions of NPS pollution in management of water quality, an aspect that is seldom considered in South Africa. The Water Research Commission (WRC) has addressed this gap by funding a number of studies, which, amongst others, have generated a knowledge review of modelling agricultural NPS pollution at multiple scales and developed an integrated modelling approach to prediction of agricultural NPS pollution. As a follow on to the latter project, the WRC is currently funding a more detailed study on the fate and transport of nutrients in agricultural catchments.

With respect to pesticides, Burger and Nel (2008) produced a scoping study report on the impact of pesticides with endocrine disrupting properties on water resources of South Africa. This led to the solicitation of a directed project to perform a more detailed investigation of the contamination of water resources by agricultural chemicals and their impact on human health. This study produced national maps of the spatial distribution of estimated use of over 200 pesticides and is an essential input in identifying important sources of agricultural chemicals in the country.

Furthermore, the study confirmed the presence of numerous pesticides in surface waters in three case study catchments representing different crop types (i.e. maize, sugar cane and sub-tropical fruit). Given the widespread use of pesticides in large and small-scale farming, in combination with their potential toxic and endocrine disrupting effects on human and ecological health, it is important to develop, evaluate and test modelling and monitoring approaches that enable farmers and catchment managers to identify and reduce their impacts on the environment. These approaches need to consider the nature of NPS pollution and the variety of geographical, climatic and management practices that influence the transport of pesticides in the environment.

There are essentially three types of management options suitable for reducing the pollution of water pollution by agricultural pesticides:

- 1. Reduce the amount of pesticide applied through:
	- a) Improving application efficiency; and
	- b) Using non-chemical (Integrated Pest Management) control measures.
- 2. Substitute current use pesticides with less toxic, less persistent or less mobile pesticides.
- 3. Reduce the physical transport of applied pesticides from fields to aquatic systems.

AIMS & OBJECTIVES

Considering the increasing challenges related to water quality management in South Africa and acknowledging the role of NPS pesticide pollution in the deterioration of water quality, the aim of this project is to improve the management of NPS pesticide pollution in South Africa through investigating and developing monitoring, modelling and risk assessment approaches that:

- 1. Identify target areas (or hotspots) where agricultural NPS pollution of pesticides is of greatest concern;
- 2. Identify specific pesticides in hotspot areas which are likely to pose the greatest risk to freshwater resources;
- 3. Identify important transport routes of contamination and reliably predict pesticide concentrations derived from these transport routes;
- 4. Provide guidance on reducing impacts through:
	- a) Improving application, or
	- b) Selecting pesticides which pose less risk to the aquatic environment, or
	- c) Reducing the transport of chemicals from source to the aquatic environment.

These aims were addressed through addressing the following project objectives:

- 1. Knowledge review of current management practices aimed at reducing pesticide impacts in the environment.
- 2. Identify priority areas for implementation of management practices through integrating recently developed pesticide use maps with geographical data to produce maps of aggregated risk to the environment.
- 3. Develop calibration guidelines for small-scale farmers designed to reduce pesticide application rates and associated environmental impacts.
- 4. Develop relative risk-based guidelines that enable farmers to identify the relative environmental risk of different pesticides registered for use on a crop.
- 5. Perform field and catchment scale modelling studies under different cropping systems to characterise pesticide contamination in associated water resources.
- 6. Perform field and catchment scale monitoring studies under different cropping systems to test novel monitoring techniques and validate the accuracy of model outputs in predicting pesticide transport.

This report documents the research conducted as part of Objectives 2 and 4. Research outputs associated with Objectives 1, 3, 5 and 6 have been described in Volume 1 of this report.

METHODOLOGY

Updated pesticide use maps

The development of pesticide risk maps for South Africa builds on the pesticide use maps produced in WRC Project No. K5/1956. Maps produced previously were based on 2009 data. Pesticide maps were therefore updated using more recent pesticide use data for South Africa (2014) which was purchased from GfK Kynetec, an international market research company. Data was provided as the aggregated total amount of active ingredient applied to all crops as well as the disaggregated application per crop.

The spatial distribution of pesticide use is dependent on the spatial distribution of the crops to which they are applied. As for the development of maps based on 2009 pesticide use data, the Census of Agriculture Provincial Statistics performed by Statistics South Africa in 2002 was used to estimate the spatial distribution of crop production in South Africa. Data from a more recent survey conducted in 2007 was scrutinised but was found to lack sufficient detail regarding the number of different crops included in the survey.

The 2002 census collected data on crop area (ha) and production (tonnes) for commercial crops at a magisterial district level, which was used to estimate the percentage agricultural area covered by a specific crop type within a magisterial district. It is important to note that the magisterial district boundaries in South Africa have changed since 2002. For the purposes of this project, the magisterial district boundaries as demarcated in 2002 were used for the spatial mapping of crop coverage and pesticide use. Pesticide use, land cover, and geographical data need to be represented in hydrologically distinct zones (e.g. catchments) in order to predict risk. Pesticide use quantified at a magisterial district scale was therefore normalised to a quaternary catchment scale resulting in maps displaying the estimated application rate (kg/ha) of each pesticide per quaternary catchment.

Pesticide risk maps

National-level risk mapping was undertaken to identify specific catchments within South Africa with the greatest potential for impacts on aquatic biodiversity from estimated agricultural use of pesticides. This a development on the maps of pesticide use produced in WRC Project No. K5/1956 in that the method aggregates the risk of all pesticides applied on all crops within a catchment to provide a more refined spatial overview of hotspot areas. This is an important step within the general approach adopted in the larger project framework, which is to identify priority areas where the implementation of management practices is most urgently required.

Pesticide use maps, length of river network, geographical variables (slope, soil and rainfall) and physicochemical (half-life and Koc) data were integrated in a geographical information system (GIS) to make first order predictions of pesticide concentrations in the river network of each quaternary catchment in South Africa. These concentrations were compared with ecotoxicological data (algae, insects and fish) for each pesticide to determine the number of toxic units of the pesticide present in the system (toxic unit is the ratio of pesticide concentration to the specific toxicity value for a given pesticide). This was done for every pesticide (for which use data was available) in every quaternary catchment.

The total risk per catchment was determined by summing the estimated toxic units for every pesticide applied in a catchment and is therefore representative of the aggregated risk posed by all pesticides applied in the catchment. This utility enables users to identify hotspot areas/catchments based on the ability of pesticides to move into water resources and/or potential risk to aquatic ecosystem. As pesticide use is directly related to crop type, it is possible to identify the contribution of different crops to the total aggregated risk. The selection of priority areas for management interventions can therefore be informed by the total toxic unit score for the catchment as well as important crops that contribute to the score. This information can be used in targeting monitoring and mitigation campaigns designed to protect the aquatic environment.

Pesticide Risk Indicator

Prioritisation of pesticide risks (i.e. for designing monitoring programmes) or the substitution of pesticides requires the need for a comparative assessment of different pesticides that can be potentially applied to a pest on a crop. This approach is relevant given that several pesticides may be registered for use against a specific pest on a specific crop, each of which will vary in terms of its application rate, mobility (i.e. physicochemical properties) and toxicity to non-target biota. Environmental pesticide risk indicators provide a relative indication of the impact of pesticides, with an emphasis on relatively few and simple data input parameters.

An output of WRC Project No. K5/1956 was the development of a prioritisation tool that allowed for the prioritisation of pesticides based on their quantity of use, environmental mobility and risk to human health. The tool was designed to identify priority pesticides at a very coarse, national level. The overall aim of this current project required a more refined catchment- or farm-level approach and is focussed on aquatic ecological health as opposed to human health.

The aim of the risk-based indicator developed in this project is to:

- 1. Assist in reducing the impacts of agricultural chemicals in water resources by allowing farmers and catchment managers to easily assess the risk of different pesticides to the environment and therefore choose pesticides that pose the least risk to the environment.
- 2. Assist catchment managers in prioritising high-risk pesticides within a catchment based on the crops produced in the catchment, with a view to informing monitoring programmes or management interventions.

The approach adopted in the development of the indicator involves the development of an Excel based software tool that provides an indication of the relative risk of pesticides applied to a crop. The tool enables the user to select a crop, for which a list of registered active ingredients is automatically generated. Based on the input of several easily obtainable input parameters, the tool provides a relative assessment of the risk of each pesticide to aquatic ecosystem health. This risk is estimated by calculating a Predicted Environmental Concentration (PEC) for each pesticide, which is compared to a toxicity concentration for each pesticide. The ratio of the PEC to the toxicity value, gives an indication of the risk of the pesticide and comparison of ratios for all pesticides provides a means to compare risks between pesticides.

The calculation of the PEC assumes contribution of pesticide loads derived from runoff and spray drift. The load of pesticide derived from each transport pathway is calculated by the spreadsheet, using simple, empirical formulae. These loads are expressed as a concentration through the input of basic stream dimension and flow parameters.

The spreadsheet relies on three databases, which have been developed as part of this project and include the following:

- A database of active ingredients registered for use on specific pests for specific crops;
- A pesticide physicochemical database;
- A pesticide toxicity database;
- A database of crop-specific pesticide application rates.

RESULTS & DISCUSSION

Updated pesticide use maps

The maps produced here rely on more recent pesticide use data and therefore represent an update of the maps produced in WRC Project No. K5/1956. Excel databases of estimated pesticide application (kg) and pesticide application rates (kg/ha) for more than 200 active ingredients for all quaternary catchments in South Africa have been produced. These databases have been incorporated into a GIS shapefile of quaternary catchments for South Africa from which maps illustrating the application of each pesticide in each quaternary catchment can be produced. The maps provide important information not only in terms of estimated application rates but also in terms of identifying where in the country specific pesticides are most likely being applied. As pesticide application has been quantified according to hydrologically distinct units, it is possible to apply catchment scale modelling approaches to estimate pesticide transport and associated risk per quaternary catchment.

As with the development of the first version of the maps, it is important to note the limitations associated with the assumptions used in the production of the maps. These include the following:

1. The magisterial district coverage is based on the 2002 Census of Agriculture Provincial Statistics and did not represent total coverage, as accurate statistics were dependent on farmers that responded to the census. Data was therefore normalised to reflect actual crop coverage as reported by the FAO (i.e. the area of each crop type in a magisterial district was multiplied by the ratio of total national area reported by STATSSA to total national area reported by the FAO);

- 2. The methodology assumes that a specific pesticide was evenly distributed to a specific crop regardless of the magisterial district the crop was produced in. Pesticide use data as displayed in the maps may therefore not reflect the local variability of pesticide management practices found within a magisterial district;
- 3. Because the agricultural land cover does not discriminate between different crop types, pesticide use was aggregated up to a magisterial district level and assumed to be distributed across all agricultural land within a magisterial district. All agricultural land cover that fell within a magisterial district was therefore assigned a pesticide use category for the pesticide in question;
- 4. Crop production statistics may not have been available for all magisterial districts where a pesticide may have been applied to agricultural land, and therefore, are not displayed on the maps; and
- 5. Pesticide use estimates are based on market research data for the year 2014

Pesticide risk maps

Maps illustrating the relative risk of pesticide application per quaternary catchment to algae, invertebrates and fish were developed. In general, risks are highest for the following areas:

- Berg, Breede and Olifants primary catchments in the south-western Cape
- Along the lower reaches of the Orange River in the vicinity of Upington in the Lower Orange primary catchment
- Upper and Middle Vaal primary catchments
- The eastern sections of the Luvulvhu and Letaba and Inkomati primary catchments
- The Usuthu to Mhlatuze, Tugela and Myoti to Umzimkulu primary catchments.
- Quaternary catchments located in the southern section of the Fish to Tsitsikamma primary catchment

The spatial extent of risks generally decreases in the order from algae to invertebrates to fish. This is a function of the high quantities of herbicides (and associated increased toxicity to algae) used on South African crops in comparison to fungicides and insecticides, which are generally less toxic towards algae. The maps provide an initial assessment in identifying priority areas where pesticide use may be impacting on aquatic ecosystem health. Databases produced as part of this deliverable can be used to identify further, which specific pesticides contribute most towards the Toxic Unit value for a quaternary catchment of interest. For example, pesticide application in quaternary catchments G10D, G21C and G21E (Berg River primary catchment) is highlighted as being very high risk towards fish. This is of concern given the high prevalence of endemic, endangered fish species that generally occur in the south-western Cape. Interrogation of the toxic unit database indicates that for quaternary catchment G10D (as an example), three pesticides (gamma-cyhalothrin, chlorpyrifos and esfenvalerate) clearly contribute significantly towards the total Toxic Unit score for the catchment (77.9% of the total).

Risk indicator

Two versions of the indicator have been developed. The Generic Risk Indicator can be used when the user has some knowledge of which crops are produced in a catchment but does not have any information of what specific pesticides are applied to the crop. The Specific Risk Indicator can be used when the user has

some knowledge of which pesticides are applied to a particular crop in a catchment (including their application rate) or can also be used to assess the relative risk of several specific pesticides of interest.

The Data Input Tab for the Generic Indicator allows the user to add relevant geographical information required for the calculation of the PEC and the associated RQ. These input parameters are easy to obtain using readily available resources (e.g. Google Earth, Cape Farm Mapper, etc.) or GIS data. Based on the input of parameters in this tab, the PEC for all pesticides that are likely to be used on the crop is automatically calculated through reference to lookup pesticide use tables (as supplied by the GfK Kynetec database) included in the spreadsheet. The Generic Indicator therefore provides a list of pesticides (and associated risks) that are likely to be applied on a crop (as determined by market research).

As mentioned above the Specific Risk Indicator can be used when the user has some knowledge of which pesticides are applied to a particular crop in a catchment or can also be used to assess the relative risk of several specific pesticides of interest. In addition to the catchment data input that is required for the Generic Risk Indicator an additional input box is provided where specific pesticides can be selected (up to ten active ingredients can be selected) and the application rate can be included.

CONCLUSION

The combination of resources produced as part of this deliverable are clearly an effective means of:

- 1. Identifying and/or prioritising catchments where agricultural pesticide use may pose an unacceptable risk to aquatic biota.
- 2. Identifying specific pesticides that contribute the greatest potential risk to aquatic biota.
- 3. Designing appropriate monitoring and management programmes aimed at mitigating risks in the catchment area.

The outputs of the Generic Pesticide Risk Indicator are based on crop-specific application rates obtained from the pesticide use database as provided in the Appendix to this report. The utility of the outputs can be summarised as follows:

- Provides a best guess of which pesticides could be applied in a catchment area where no detailed pesticide application information is available;
- Provides an indication of the relative risks these pesticides may pose to algae, invertebrates and fish;
- The outputs differentiate risks associated with runoff and spray drift and therefore provide guidance on monitoring strategies as well as which management interventions are most likely needed to control the source of pesticides in the catchment (i.e. spray drift will require different management interventions in comparison to runoff).
- The outputs can identify which pesticides are likely to be used in high quantities and the risk associated with these pesticides and can therefore be used to identify target pesticides for further monitoring/screening/management;
- The indicator can be used to assess how changes in certain crop management inputs (i.e. increased buffer widths, vegetated buffer zones, etc.) can potentially affect the risk associated with the use of a pesticide; and
- The indicator can be applied in different sub-catchments or tributaries in order to identify specific hotspot areas within a larger catchment area.

The outputs of the Specific Pesticide Risk Indicator are based on user defined pesticides and associated application rates. The utility of the outputs can be summarised as follows:

- Provides a higher confidence estimate of relative risks based on more detailed knowledge of pesticide application programmes in the catchment;
- The outputs differentiate risks associated with runoff and spray drift and therefore provide guidance on monitoring strategies as well as which management interventions are most likely needed to control the source of pesticides in the catchment (i.e. spray drift will require different management interventions in comparison to runoff);
- The indicator can be applied in different sub-catchments or tributaries in order to identify specific hotspot areas within a larger catchment area;
- The indicator can be used to assess how changes in certain crop management inputs (i.e. increased buffer widths, vegetated buffer zones, etc.) can potentially affect the risk associated with the use of a pesticide; and
- The indicator can be used to compare specific pesticides of interest with a view to identifying lower risk pesticides (i.e. for application to crops) or higher risk pesticides that might need to be included in a monitoring programme.

TECHNOLOGY TRANSFER & INNOVATION

South Africa has a well-developed agricultural sector, which uses high quantities of numerous different pesticides on an annual basis. We have very little information on the impact that this pesticide use has on our very limited freshwater resources and no reliable resources are available for identifying risks of pesticides to water resources in South Africa. The outputs of the project have developed pesticide use and risk maps and a risk indicator and have tested software models aimed at prioritising risks of pesticides and predicting environmental concentrations in water resources. These resources significantly improve our ability to assess risks of pesticides to the aquatic ecosystem in South Africa. While similar products have been developed throughout the rest of the world, no such products have been developed for South Africa. The outputs of this project therefore help improve our understanding of the potential impacts of pesticides at multiple scales (i.e. field, catchment and national scale) and directly addresses the Department of Agriculture Draft Pesticide Management Policy (2006) in which the need to minimise hazards and risks to health and the environment was highlighted.

Risk maps

The following resources have been produced in the development of pesticide risk maps for South Africa:

Risk indicator

The following Excel-based pesticide risk indicators have been produced as part of this project:

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This project was initiated by the Water Research Commission (WRC) of South Africa through a nonsolicited call. The Freshwater Research Centre (FRC) together with collaborators from the University of Pretoria (UP) and the Agricultural Research Council (ARC) submitted a proposal and carried out the research to meet the objectives stipulated in proposal. The project was managed by the WRC.

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CHAPTER 1: DEVELOPMENT OF PESTICIDE RISK MAPS FOR AQUATIC ECOSYSTEMS IN SOUTH AFRICA

By:

James M. Dabrowski (Freshwater Research Centre)

1.1 INTRODUCTION

Given the large spatial footprint associated with nonpoint source (NPS) pollution (many diffuse sources covering a large area) and the large number of active ingredients registered for use in South Africa (over 200) a key step in managing agricultural pesticides is to identify areas where the risks associated with pesticide use are likely to be high (hotspots) (Figure 1-1). While monitoring data is ideal for identifying such hotspot areas, this is generally lacking in South Africa. Hence, the use of surrogate indicators of contamination is important to identify areas where catchment specific monitoring and management programmes can be implemented. Such indicators need to provide some indication of potential risk of pesticide use to the aquatic environment.

According to the framework presented in Figure 1-1 this step can be regarded as a screening approach and makes use of a number of readily available spatial, toxicity and physicochemical databases to make a relative assessment of the potential risks posed by pesticides to aquatic ecosystems. Once target areas and priority pesticides have been identified, a more detailed assessment can be conducted with the ultimate aim being to minimise risks to the environment.

Figure 1-1: Nonpoint source pesticide assessment framework.

WRC Project K5/1956 produced maps providing estimates of application rates (kg.ha-1) of over 200 different active ingredients registered for use in South Africa (Dabrowski, 2015). The designed maps are intended to provide a spatial overview of the relative application rates of specific active ingredients based on the distribution of crops throughout the country. From a practical perspective, these maps are useful for identifying hotspot areas if the risk associated with the use of a specific pesticide is already known. For example, there may be a need to manage atrazine levels in water resources due to a known

human or ecological health risk associated with exposure to atrazine. In this instance, the maps could be used to identify areas where atrazine is applied in high quantities to prioritise further management interventions.

It is however important to note that, while an important contributing factor, quantity of use alone does not necessarily translate into risk. Risk occurs due to a combination of factors, including the quantity of use, the likelihood of transport from the point of use into a receiving water body and ultimately exposure to a concentration that may or may not exceed a specific toxicity threshold. Pesticide transport is influenced by several factors including climatic and geographic conditions and physicochemical properties of pesticides (which vary significantly from pesticide to another). Therefore, while quantity of use provides a good indication of where risk may occur, these other factors, when taken into consideration, provide a more refined estimate of the likelihood that a risk will in fact occur.

From the perspective of identifying high-risk areas, the pesticide use maps provide very important information not only in terms of estimated application rates, but also in terms of identifying where in the country specific pesticides are most likely being applied. Integrating this information together with spatial (e.g. soil type, slope, rainfall, etc.) and toxicity data using an environmental risk mapping technique can identify hotspot areas at a national scale.

Environmental risk mapping is a technique whereby a Geographic Information System (GIS) is used to superimpose sets of spatial information (i.e. those that influence transport) to combine variability in terms of pollution pressure and vulnerability (Brown *et al.* 2007). Information regarding the spatial distribution of risk to surface waters arising from pesticide use can be used to target monitoring campaigns, to identify priority areas for management interventions, and to investigate optimal mitigation strategies.

1.2 UPDATING PESTICIDE USE DATA (2014)

The development of pesticide risk maps for South Africa builds on the pesticide use maps produced in WRC Project No. K5/1956 (Dabrowski, 2015). Maps produced previously were based on 2009 data. Pesticide data was updated using more recent data (2014) using the methodology described in Dabrowski (2015). Pesticide use data for South Africa was purchased from GfK Kynetec (http://www.gfk.com/gfk-kynetec/), an international market research company. Data was provided as the aggregated total amount of active ingredient applied to all crops as well as the disaggregated application per crop.

In comparison to data for 2009, the 2014 database showed the following differences:

- Stone fruit were grouped as a single crop class (including apricots and plums)
- Sub-tropical fruit were grouped as a single class (including avocadoes, bananas, mangoes and papayas)
- Beans were grouped as a single crop class (including dry and green beans)

Additional crops for which data was obtained included the following:

- Lucerne
- Vegetables (grouped as a single crop class)

Apples	Pears
Barley	Pineapples
Beans	Potatoes
Citrus	Sorghum
Cotton	Soya beans
Grapes-table	Stone fruit
Grapes-wine	Sub-tropical Fruit
Groundnuts	Sugar cane
Lucerne	Sunflower
Maize	Tomatoes
Peaches	Vegetables
	Wheat

Table 1-1: Crops included in the 2014 GfK Database

Pesticide data for 2014 is included in the Appendix to this report.

1.2.1 Crop Distribution

The spatial distribution of pesticide use is dependent on the spatial distribution of the crops to which they are applied. As for the development of maps based on 2009 pesticide use data, the Census of Agriculture Provincial Statistics performed by Statistics South Africa in 2002 (STATSSA, 2002) was used to estimate the spatial distribution of crop production in South Africa. Data from a more recent survey conducted in 2007 was scrutinised but was found to lack sufficient detail regarding the number of different crops included in the survey.

The 2002 census collected data on crop area (ha) and production (tonnes) for commercial crops at a magisterial district level, which was used to estimate the percentage agricultural area covered by a specific crop type within a magisterial district. It is important to note that the magisterial district boundaries in South Africa have changed since 2002. For the purposes of this project, the magisterial district boundaries as demarcated in 2002 were used for the spatial mapping of crop coverage and pesticide use. Furthermore, the census provides data only for farmers that responded. The census therefore underestimates the total area and production of a crop at a magisterial district and national level, but does provide as accurate an estimate of the relative distribution of crop coverage and production as is possible at this level of spatial detail.

The census data was therefore normalised to consider this under-estimation, as well as to account for changes in area and production over time to provide an estimate of total crop coverage per magisterial district in 2014. The normalisation procedure compared total crop coverage estimated by the 2002 STATSSA census data (i.e. the sum of the area of each crop type for all nine provinces) with total crop area statistics collected by the United Nations Food and Agriculture Organisation (FAO) for the year 2014 (FAOSTAT, 2016). The crop normalisation quotient was expressed as the ratio of FAO to STATSSA crop coverage. The crop coverage reported by STATSSA for each magisterial district in South Africa was multiplied by the respective crop normalisation quotient to derive a normalised crop coverage for each crop type in each magisterial district. The normalised area of each crop in each magisterial district was expressed as a percentage of the national crop area for the crop.

Table 1-2: National crop area statistics reported by STATSSA and the FAO. The ratio of FAO to STATSSA statistics was used to normalise crop area data for each magisterial district in South Africa.

Once crop coverage data had been normalised, the methodology described in detail in Dabrowski (2015) was used to quantify pesticide application (kg) and pesticide application rates (kg.ha-1) on agricultural land in each magisterial district for the year 2014.

The amount of each pesticide applied to a crop was expressed as a percentage of the total national application. For example, approximately 34% of the total national of azinphos-methyl is applied to maize (Table 1-3). These percentages were used to estimate the percentage of the total amount of each pesticide applied to each crop in each magisterial district *(P%)*:

$$
P\%_{x,y,z}=\frac{Area_{y,z}}{100}\times CAPp_x
$$

Where *Area* is the proportion of crop type *(y)* in a magisterial district *(z)* expressed as a percentage of the total national coverage of the crop and *CApp* is the proportion of the pesticide *(x)* applied to the crop, expressed as a percentage of the total application of the pesticide. This assumes that a specific pesticide was applied equally (or at an identical application rate) to a specific crop regardless of the magisterial district the crop was produced in.

Table 1-3: Example of a summary of the application of azinphos-methyl to crops produced in South Africa

The total estimated quantity (Pq , in kg) of each applied pesticide (x) to each crop type (y) in each magisterial district *(z)* was calculated as follows:

$$
Pq_{x,y,z} = \frac{P\%_{x,y,z}}{100} \times TApp_x \, ;
$$

Where *TApp* is the total quantity of pesticide *x* applied to all crops in the country. From this data, it was possible to estimate the total quantity (*Ptot*, in kg) of pesticide *(x)* applied within a magisterial district *(z)* regardless of crop type *(y)*:

$$
Ptot_{x,z} = \sum_{y=1}^{n} Pq_{x,y}
$$

1.2.2 Defining pesticide use according to quaternary catchment

Pesticide use, land cover, and geographical data need to be represented in hydrologically distinct zones (e.g. catchments) in order to predict risk. Pesticide use quantified at a magisterial district scale therefore had to be normalised to a quaternary catchment scale.

GIS was used to execute the following steps:

- 1. The latest agricultural land cover data set was used to quantify the amount of agricultural land cover (ALC) per magisterial district.
- 2. The magisterial district was intersected with the quaternary catchment layer to produce a shapefile with polygons representative of each magisterial district and quaternary catchment combination.
- 3. The ALC within each combination of magisterial district and quaternary catchment polygon was quantified using zonal statistics.

Agricultural land cover (*ALC*) in each quaternary catchment (*v*) was expressed as a percentage of the total agricultural land cover contained within the corresponding magisterial district (*z*) (e.g. quaternary catchment *v* contains 5% of agricultural land cover contained within magisterial district *z*):

$$
ALC\%_{v} = \frac{ALC_{v,z}}{ALC_{z}} \times 100 ;
$$

The quantity of each pesticide *(x)* applied in each combination of magisterial district (*z*) and quaternary catchment (*v*) combination was calculated by multiplying the percentage of agricultural land cover in the quaternary catchment (*ALC%v*) by the total amount (*Ptot*) of pesticide (*x*) applied in the magisterial district (*z*):

$$
P_{v,x,z} = \frac{ALC\%_{v}}{100} \times Ptot_{x,z}
$$

From this data, it was possible to estimate the total quantity (*Ptot*, in kg) of pesticide *(x)* applied within a quaternary catchment *(v)* by summing the application in each magisterial district falling within the quaternary catchment:

$$
Ptot_{x,v} = \sum_{z=1}^{n} P_{x,z}
$$

1.2.3 2014 Pesticide Maps for South Africa

The maps produced here rely on more recent pesticide use data and therefore represent an update of the maps produced in WRC Project No. K5/1956. Excel databases of estimated pesticide application (kg) and pesticide application rates (kg/ha) for more than 200 active ingredients for all quaternary catchments in South Africa have been produced. These databases have been incorporated into a GIS shapefile of quaternary catchments for South Africa from which maps illustrating the application of each pesticide in each quaternary catchment can be produced (Figure 1-2). The maps provide important information not only in terms of estimated application rates but also in terms of identifying where in the country specific pesticides are most likely being applied. As pesticide application has been quantified according to hydrologically distinct units, it is possible to apply catchment scale modelling approaches to estimate pesticide transport and associated risk per quaternary catchment.

As with the development of the first version of the maps, it is important to note the limitations associated with the assumptions used in the production of the maps. These include the following:

- 1. The magisterial district coverage is based on the 2002 Census of Agriculture Provincial Statistics and did not represent all total coverage, as accurate statistics were dependent on farmers that responded to the census. Data was therefore normalised to reflect actual crop coverage as reported by the FAO (i.e. the area of each crop type in a magisterial district was multiplied by the ratio of total national area reported by STATSSA to total national area reported by the FAO);
- 2. The methodology assumes that a specific pesticide was evenly distributed to a specific crop regardless of the magisterial district the crop was produced in. Pesticide use data as displayed in the maps may therefore not reflect the local variability of pesticide management practices found within a magisterial district;
- 3. Because the agricultural land cover does not discriminate between different crop types, pesticide use was aggregated up to a magisterial district level and assumed to be distributed across all agricultural land within a magisterial district. All agricultural land cover that fell within a magisterial district was therefore assigned a pesticide use category for the pesticide in question;
- 4. Crop production statistics may not have been available for all magisterial districts where a pesticide may have been applied to agricultural land, and therefore, are not displayed on the maps; and

5. Pesticide use estimates are based on market research data for the year 2014.

Figure 1-2: Example of map displaying estimated crop application rates of atrazine per quaternary catchment in South Africa.

1.3 PESTICIDE RISK MAPS

1.3.1 General approach

The development of pesticide risk maps for South Africa builds on the pesticide use maps produced in WRC Project No. K5/1956 (Dabrowski, 2015). National-level risk mapping was undertaken to identify specific catchments within South Africa with the greatest potential for impacts on aquatic biodiversity from normal agricultural use of pesticides. This a development on the maps of pesticide use produced in WRC Project No. K5/1956 in that the method aggregates the risk of all pesticides applied on all crops within a catchment to provide a more refined spatial overview of hotspot areas. This is an important step within the general approach adopted in the larger project framework, which is to initially identify priority areas where the implementation of management practices are most urgently required (Figure 1-1). Pesticide use as indicated by the pesticide use maps, length of river network, geographical variables (slope, soil and rainfall) and physicochemical (half-life and *KOC*) data will be integrated in a GIS to make first order predictions of pesticide concentrations in the river network of each quaternary catchment in South Africa. These concentrations will be compared with ecotoxicological data (algae, insects and fish) for each pesticide to determine the number of toxic units of the pesticide present in the system (toxic unit is the ratio of pesticide concentration to the specific toxicity value for a given pesticide). This will be done for every pesticide (for which use data is available) in every quaternary catchment. The total risk per catchment can be determined by summing the estimated toxic units for every pesticide applied in a catchment and is therefore representative of the aggregated risk posed by all pesticides applied in the catchment. This utility will enable users to identify hotspot areas/catchments based on the ability of pesticides to move into water resources and/or potential risk to aquatic ecosystem. As pesticide use is directly related to crop type, it will be possible to identify the contribution of different crops to the total aggregated risk. The selection of priority areas for management interventions can therefore be informed by the total toxic unit score for the catchment as well as important crops that contribute to the score. This information can be used in targeting monitoring and mitigation campaigns designed to protect the aquatic environment.

1.3.2 OECD runoff model

The Organisation for Economic Co-operation and Development (OECD) as part of their development of a pesticide risk indicator for both human health and the environment developed a simple model designed to estimate the percentage of an applied pesticide that leaves an agricultural field as runoff following a rainfall event (OECD, 1998).

1.3.2.1 Application in South Africa

The OECD model has been rigorously tested in the Lourens River catchment, Western Cape, South Africa. Dabrowski and Schulz (2003) used the model to estimate loadings and Predicted Environmental Concentrations (PECs) of azinphos-methyl in the main stem and tributaries of the Lourens River. PECs corresponded well with measured concentrations taken during monitored runoff events and demonstrates the potential of the model for use in risk assessment for registration purposes. An additional study conducted in the Lourens River catchment used the equation as an indicator to predict the relative mobility and occurrence of several pesticides in the river following runoff events (Dabrowski and Balderacchi, 2013). Samples were collected weekly at five sites from the beginning of the spraying season (October) until the beginning of the rainy season (April) and were semi-quantitatively analysed for relevant pesticides applied according to the local farmers spraying programme.

A comparison of indicator outputs to monitoring data showed that the OECD model could:

- 1. Differentiate between highly contaminated and less contaminated sites;
- 2. Accurately predict the relative exposure potential of specific pesticides across and within sites; and
- 3. Identify the most important route of transport for all pesticides at all sites.

In contrast to the Dabrowski and Schulz (2003) study, the aim of this study was not to estimate PECs but rather to provide a relative indication of exposure and associated risk. In this context, the model proved to be a reliable screening tool and could therefore be applicable in a lower tier assessment. Another study conducted in the Lourens River catchment examined the effect of erosion rills on the efficiency of vegetative buffer strips (Stehle *et al.*, 2016). The results showed that erosion rills are common in buffer strips adjacent to tributaries and represent concentrated entry pathways of pesticide runoff into the tributaries during rainfall events. Exposure modelling using the OECD runoff equation showed that measured pesticide surface water concentrations correlated significantly with runoff losses predicted by model scenarios in which buffer strip width was set to zero at sites with erosion rills. In contrast, no relationship between predicted runoff losses and in-stream pesticide concentrations were detected in the model scenario that neglected erosion rills and thus assumed efficient buffer strips.

Application of the OECD model in this context was able to show that erosion rills may substantially reduce buffer strip pesticide retention efficacies during runoff events. This suggests that the capability of buffer strips as a risk mitigation tool for runoff is largely over-estimated in current regulatory risk assessment procedures conducted for pesticide authorisation.

In summary, results from the above-mentioned studies indicate that the OECD model can be reliably used to estimate pesticide concentrations in runoff in South African conditions.

1.3.2.2 Application internationally

The Pesticide Impact Rating Index (PIRI) developed by CSIRO, Australia is an example of a risk indicator that is regularly applied in managing pesticides in agricultural catchments (Kookana *et al.* 2005). The index integrates built in toxicity and physicochemical data with user inputs of application data (i.e. application rate, frequency of application) and readily available geographical data (e.g. soil type, rainfall amount, slope, etc.) to provide a relative index of a mobility and risk for a number of pesticides. Runoff transport is estimated using the OECD runoff formula by Reus *et al.* (1999).

Figure 1-3 provides a visual representation of the relative mobility and risk of herbicides applied to sugarcane. By entering geographical data specific to several different catchments (e.g. soil type, slope, etc.) the relative risk of different pesticides applied across the different catchments can also be determined. In addition to the PIRI model, the Australian Pesticides and Veterinary Medicines Authority (APVMA) uses the OECD runoff model in the risk assessment process for the registration of pesticides in Australia (Lee-Steere, 2007).

Figure 1-3: Example of output of the PIRI risk indicator showing a comparison of the relative mobility (left) and risk (right) of different pesticides applied to a crop (Kookana et al. 2005)..

1.3.3 Estimating pesticide loss

The equation essentially consists of empirical and physical components, including a hydrological model predicting runoff amounts, catchment related factors, which influence the extent of runoff, a first-order kinetic model describing the degradation of a pesticide and a term referring to the proportion of pesticide occurring in the water phase of runoff.

The equation is as follows:

$$
L\%_{runoff} = \frac{Q}{P} \times f \times \exp(-3 \times \frac{ln2}{DT_{50\, soil}}) \times \frac{100}{(1+Kd)}
$$

Where

L%runoff = percentage of application dose being available in runoff water as a dissolved substance; *Q* = runoff amount (mm) calculated according to hydrological models (Lutz (1984) and Maniak (1992)); *P* = precipitation amount (mm);

DT50soil= half-life of active ingredient in soil (d);

 $f = f1 \times f2 \times f3$, the correction factor reflecting the influence of slope ($f1 = 0.02153 \times slope + 0.001423 \times s$) *slope2*), plant interception (*PI*), the percentage of applied pesticide intercepted by trees in the orchards (*f2 = 1 - PI/100*), and buffer width (*f3 = 0.83WBZ*, and WBZ is the width of buffer zone [m]; if the buffer zone is not densely covered with plants, the width is set to zero);

 $t =$ time (d) between application and rainfall;

Kd = (Koc x %OC), a factor reflecting the tendency of the pesticide to bind to organic carbon in the soil, where *Koc* is the sorption coefficient of the active ingredient to organic carbon (mL/g) and *OC%* is the organic carbon content of the soil.

Tables developed by Lutz (1984) and Maniak (1992) are used to obtain the Q value corresponding to rainfall events above 10 mm (Table 1-4). The methodology used to derive Q in this equation is relevant to German conditions. Other approaches could be used to estimate runoff amounts (e.g. the Soil Conservation Service Curve Number approach). While this is a relatively simple model, the Australian Pesticides and Veterinary Medicines Authority (APVMA) use this model for registration of pesticides in Australia (Lee-Steere, 2007).

Rainfall (mm)	1 Sandy Soil	Loamy Soil
6	0.1	0.45
8	0.28	0.82
$\overline{10}$	0.54	1.29
12	0.88	1.86
14	1.29	2.51
16	1.78	3.24
18	2.32	4.05
20	2.92	4.93
$\overline{22}$	3.58	5.88
24	4.29	6.88
26	5.04	7.95
$\overline{28}$	5.84	9.06
30	6.69	10.23
32	7.57	11.45
34	8.48	12.7
36	9.44	14
38	10.42	15.34
40	11.43	16.71
42	12.47	18.11
44	13.53	19.54
46	14.62	21.01
48	15.73	22.5
50	16.87	24.01

Table 1-4: Runoff volumes (mm) for sandy and loamy soils according to the model of Lutz (1984) and Maniak (1992).

1.3.4 Model input data

1.3.4.1 GIS data

Sources of all relevant GIS data are included in Table 1-5. Agricultural land cover was extracted from the 2009 National Land Cover and was used to identify the location of agricultural crops per quaternary catchment. This layer was used to extract corresponding soil type (Figure 1-4), organic carbon content (Figure 1-5) and slope data (Figure 1-6) required for the runoff calculation. The OECD runoff equation determines the total amount of runoff (mm) following a rainfall event based on whether the soil type is a sand or loam soil. The WR90 soil coverage was therefore used to identify soils according to these two categories (Table 1-6, Figure 1-4).

Table 1-5: Source of GIS data used to extract geographical information required to predict.

GIS Layer	Source
Agricultural Land Cover	2013-14 South African National Land-cover dataset
Soil Type	Derived from WR90 Soil Map
Organic Carbon Content	ARC-ISWC
Slope	30 m SRTM Digital Elevation Model

Table 1-6: WR90 soil texture classes used to define loamy and sandy soils for the purposes of predicting runoff related pesticide input using the OECD model.

Figure 1-4: Map illustrating the soil categories derived for predicting runoff related pesticide losses in South Africa.

Figure 1-5: Map illustrating the organic carbon content (percentage) of soil covered by agricultural land use in South Africa.

Figure 1-6: Map illustrating slope categories (percentage) of agricultural land in South Africa.

1.3.4.2 Pesticide physicochemical data

Physico-chemical data (half-life and *KOC*) for each pesticide was obtained from the Pesticide Properties Database (Lewis *et al.* 2016) and captured in an excel database. Toxicity data for algae, *Daphnia magna* and fish was also obtained.

1.3.5 GIS processing

All GIS processing was done using QGIS 2.18 according to the following steps:

- 1. An agricultural land cover raster was extracted from the national land cover data set.
- 2. The agricultural land cover raster was overlaid with the soil type raster to assign a soil type (e.g. loamy or sandy soil) to each agricultural land cover pixel (Figure 1-4).
- 3. The agricultural soil raster (Figure 1-4) was overlaid with the slope raster (Figure 1-6) to assign a slope percentage to each agricultural soil pixel. Zonal statistics was then used to calculate the average slope for each agricultural soil type in each quaternary catchment. This data was exported into an excel database.
- 4. The agricultural soil raster (Figure 1-4) was overlaid with the organic carbon raster (Figure 1-5) to assign an organic carbon content (percentage) to each agricultural soil pixel. Zonal statistics was then used to calculate the average organic carbon percentage for each agricultural soil type in each quaternary catchment. This data was exported into an excel database.

1.3.6 Predicted environmental load (PEL)

Pesticide application data per quaternary catchment (Section 1.2.2), physicochemical data (Section 1.3.4.2) and summarised geographical date per quaternary catchment (Section 1.3.5) were included in an excel database in which the runoff calculations were performed using the OECD formula (Section **Error! Reference source not found.**). The formula was applied for each pesticide applied to agricultural land under each soil type (i.e. sand and/or loam) in each quaternary catchment. The final output was the percentage of applied pesticide being available in runoff water (*L%runoff*). The following assumptions were made for predicting pesticide runoff losses:

- Calculations assumed a 20 mm rainfall event
- Pesticide application had taken 3 days prior to the rainfall event
- \bullet Plant interception (PI) was assumed to be 80%.

The PEL refers to the total amount of each pesticide (*x*) washed off from agricultural crops located in each quaternary catchment (*y*) into adjacent water bodies:

$$
PEL_{x,y} = \frac{L\%_{x,y}}{100} \times (PApp_{x,y} \times ALC_y)
$$

Where $L\%$ = estimated percentage of applied pesticide lost because of runoff and was calculated according to the equation in Section 1.3.3; $PApp =$ the application rate of the pesticide (kg/ha) and ALC = the total agricultural land cover (ha). The calculation of *L%* used average slope and organic carbon content values for each soil type in each quaternary catchment as described in Section 1.3.5.

PEL calculations were performed separately for loam and sandy soils under agricultural land cover in each quaternary catchment. This is because estimated runoff volumes (*Q*) vary according to each soil type (see Section 1.3.3). PELs for each soil type were then combined (i.e. summed together) to give a final PEL per quaternary catchment.

1.3.7 Estimating predicted environmental concentrations (PECs)

PECs were calculated based on the PEL and estimated hydrological discharge volumes for each quaternary catchment. The following assumptions were made for estimating discharge volume:

- Runoff volumes were estimated according to the OECD model 20 mm rainfall event
- The rainfall event was assumed to last for 1 hour
- Discharge in surface waters of the quaternary catchment were assumed to be equal to the mean hourly runoff (calculated from the mean annual runoff (MAR) for the catchment).

The PEC for each pesticide (*x*) in each quaternary catchment (*y*) was calculated as follows:

$$
PEC_{x,y} = \frac{(PEL_{x,y} \times 10^6)}{V_y}
$$

Where 10⁶ is a conversion factor to convert kg to mg and *V* is the total volume of water (in litres) in the quaternary catchment (*y*) in which the pesticide is dissolved.

The volume of water was assumed to be equal to the sum of the volume of water flowing in surface waters of the catchment (*Vstream*) and the volume of water derived from runoff (*Vrunoff*) for a period of 1 hour:

$$
V = V_{stream,y} + V_{runoff,y}
$$

V_{stream} was derived from the estimated MAR (m³) for each quaternary catchment:

$$
V_{stream,y} = \frac{(MAR_y \times 10^3)}{365 \times 24 \times 60}
$$

Where 10³ is a conversion factor to convert m³ to litres; MAR = the mean annual runoff (m³) for the quaternary catchment.

The volume of water entering water bodies in the quaternary catchment because of the rainfall event (*Vrunoff*) was calculated as follows:

Volume_{runoff,y} =
$$
(0.00493 \times ALC_{loam,y} \times 10^4 \times 10^3) + (0.00292 \times ALC_{sand,y} \times 10^4 \times 10^3)
$$

Where 0.00493 is the runoff amount (m) on loamy soils and 0.00292 is the runoff amount (m) on sandy soils; ALC is the agricultural land cover (ha); $10⁴$ is a conversion factor to convert hectares to m² and $10³$ is a conversion factor the convert $m³$ to litres.

1.3.8 Calculation of toxic units

Toxic units (TUs) were calculated for each pesticide (*x*) applied in each quaternary catchment (*y*). Toxic units were calculated for three taxonomic groups (*z* i.e. algae, invertebrates and fish) according to the following formula:

$$
TU_{x,y,z} = \frac{PEC_{x,y}}{TOX_{x,z}}
$$

Where *TOX* is the acute EC50 for algae and *D. magna* and the acute LC50 for fish (Lewis *et al.*, 2016). A single toxic unit value was calculated for each taxonomic group (*z*) per quaternary catchment (*y*) by summing the toxic units calculated for each pesticide (*x*) applied in the quaternary catchment:

$$
TU_{y,z} = \sum_{x=1}^{n} TU_{x,y}
$$

1.4 RESULTS

Maps illustrating the relative risk of pesticide application per quaternary catchment to algae, invertebrates and fish are presented in Figure 1-7, Figure 1-8 and Figure 1-9, respectively. In general, risks are highest for the following areas:

- Berg, Breede and Olifants primary catchments in the south-western Cape
- Along the lower reaches of the Orange River in the vicinity of Upington in the Lower Orange primary catchment
- Upper and Middle Vaal primary catchments
- The eastern sections of the Luvulvhu and Letaba and Inkomati primary catchments
- The Usuthu to Mhlatuze, Tugela and Mvoti to Umzimkulu primary catchments.
- Quaternary catchments located in the southern section of the Fish to Tsitsikamma primary catchment

Risks generally decrease in the order from algae to invertebrates to fish. This is a function of the high quantities of herbicides (and increased toxicity to algae) used on South African crops in comparison to fungicides and insecticides, which are generally less toxic towards algae. The maps as presented below provide an initial assessment in identifying priority areas where pesticide use may be impacting on aquatic ecosystem health. Databases produced as part of this deliverable can be used to identify further, which specific pesticides contribute most towards the Toxic Unit value for a quaternary catchment of interest. For example, pesticide application in quaternary catchments G10D, G21C and G21E (Berg River primary catchment) is highlighted as being very high risk towards fish (Figure 1-9). This is of concern given the high prevalence of endemic, endangered fish species that generally occur in the south-western Cape. Interrogation of the toxic unit database (ToxUnits.xlsl) indicates that for quaternary catchment G10D (as an example), three pesticides (gamma-cyhalothrin, chlorpyrifos and esfenvalerate) clearly contribute significantly towards the total Toxic Unit score for the catchment (77.9% of the total) (Table 1-7).

Table 1-7: The contribution of the top ten pesticides that contribute towards the total Toxic Unit score for the G10D quaternary catchment.

The combination of resources produced as part of this deliverable are clearly an effective means of:

- 1. Identifying and/or prioritising catchments where agricultural pesticide use may pose an unacceptable risk to aquatic biota.
- 2. Identifying specific pesticides that contribute the greatest potential risk to aquatic biota.
- 3. Designing appropriate monitoring and management programmes aimed at mitigating risks in the catchment area.

Figure 1-7: Maps showing the potential risk to algae following exposure to pesticides via runoff based on pesticide use data for the year 2014.

Figure 1-8: Maps showing the potential risk to Daphnia magna following exposure to pesticides via runoff based on pesticide use data for the year 2014.

Figure 1-9: Maps showing the potential risk to fish following exposure to pesticides via runoff based on pesticide use data for the year 2014.

1.4.1 Data Products

Table 1-8 provides a summary of electronic outputs produced as part of the development of pesticide risk maps for South Africa.

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CHAPTER 2: DEVELOPMENT OF PESTICIDE RISK INDICATOR TO DETERMINE THE RELATIVE RISKS OF PESTICIDES APPLIED IN AGRICULTURAL CATCHMENTS

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2.1 INTRODUCTION

Nonpoint source agricultural pollution is generally considered one of the major threats to surface water quality in rural areas (Loague *et al.*, 1998). Nutrients, sediments and pesticides potentially enter aquatic environments via runoff, leaching and spray drift and pose a risk to the communities that inhabit them. Of all nonpoint source pollutants, pesticides are among the most crucial chemical stressors, simply because of their extremely high toxicity to many non-target aquatic organisms (e.g. algae, fish and macroinvertebrates).

Numerous studies have now documented the occurrence of pesticides in South African surface waters. The most extensive research has been performed in the Lourens River, in the Western Cape of South Africa. Studies have shown that current-use insecticides frequently enter the river via runoff and spray drift (Schulz, 2001a; Schulz, 2001b; Dabrowski and Schulz, 2003). Although contamination is transient, field, experimental microcosm (Schulz *et al.*, 2002) and in-situ bioassay studies (Schulz, 2003) have shown that measured pesticide levels pose significant risks to aquatic macroinvertebrate communities in the Lourens River.

Another study measured high levels of chlorpyrifos and endosulfan in several agricultural catchments throughout the Western Cape (London *et al.*, 2000). A more recent study, where monitoring was conducted quarterly in three different catchments (Letsitele, Komati and Vals River catchments in Limpopo, Mpumalanga and Free State, respectively), each representative of different crop types (sub-tropical fruit, sugar cane and citrus, maize), showed regular contamination by current-use pesticides in all three catchments (Dabrowski, 2015).

Despite these findings, the potential impact of pesticides in South African surface waters has been a low priority and is generally not considered in aspects of water resource management, especially concerning routine monitoring. Even in developed countries, despite strict regulatory procedures, pesticides are frequently detected in surface waters at concentrations that exceed environmentally acceptable levels (Stehle and Schulz, 2015). Although intensive monitoring can identify pesticide impacts, the cost of sampling and analysis makes this a highly costly exercise (particularly in the context of less developed countries such as South Africa). There is therefore a need to develop cost effective decision support methods of predicting the environmental risks of pesticide use at a field or catchment scale, which can be used to prioritise specific pesticides for monitoring purposes and identify spatial patterns of contamination.

At a farm management level, substitution of use of high-risk pesticides with those that pose lower risk is one of several operational Best Management Practices (BMPs) aimed at reducing pesticide risks to the environment. Risk is dependent not only on the toxicity of a pesticide, but also on the quantity of use, and the mobility of the pesticide in the environment. A highly toxic pesticide used in low quantities, with low mobility for example, could pose a lower risk to the environment than a less toxic pesticide used in higher quantities with higher mobility. Ultimately, the combination of the concentration of the pesticide in the water body and its toxicity threshold will determine the level of risk, which should be used as the basis upon which to make decisions on substituting one pesticide for another.

2.2 RISK INDICATORS

Prioritisation of pesticide risks (i.e. for designing monitoring programmes) or the substitution of pesticides requires the need for a comparative assessment of different pesticides that can be potentially applied to a pest on a crop. This approach is relevant given that several pesticides may be registered for use against a specific pest on a specific crop, each of which will vary in terms of its application rate, mobility (i.e. physicochemical properties) and toxicity to non-target biota. Environmental pesticide risk indicators provide a relative indication of the impact of pesticides (Levitan *et al.*, 1995), with an emphasis on relatively few and simple data input parameters (Kookana *et al.*, 2005). Risk indicators vary greatly in terms of their purpose and methodologies, and are often very broad in scope covering, for example, the impact on aquatic organisms, soil organisms, bees, occupational exposure and human health effects (Reus *et al.*, 2002). With respect to aquatic ecosystems, risk indicators are generally regarded as lower tier risk assessment tools that provide a relative assessment of the environmental impact of pesticides through integration of multiple catchment-scale factors that influence the fate, transport and toxicity of pesticides. These factors include site specific geographical conditions (e.g. soil type, soil organic matter content, water input, slope of land, soil loss, recharge rate, depth of water table etc), ecotoxicological (e.g. LC50 data or species sensitivity distributions) and environmental fate (e.g. half-life and *KOC* values) properties and pesticide use data (Sánchez-Bayo *et al.*, 2002; Sala and Vighi, 2008). Empirical pesticide transport equations integrate these factors and calculate pesticide loading and associated PECs via different routes of pesticide transport (i.e. spray drift and runoff). By comparing PECs to a toxicity endpoint for each pesticide (e.g. acute LC50 for aquatic biota), the relative risk of different pesticides can be made. The objective in such an approach is to make a relative estimate of the differences in risk associated with the use of the different pesticides. This can allow a farmer or catchment manager to identify which of several different pesticides pose the lowest risk to the environment. Similarly, water resource/catchment managers could use outputs to identify which high risk pesticides could be potentially used in a catchment based on the type of crops produced in the catchment.

2.3 AIMS & OBJECTIVES

An output of WRC Project No. K5/1956 was the development of a prioritisation tool that allowed for the prioritisation of pesticides based on their quantity of use, environmental mobility and risk to human health (Dabrowski et al, 2014). The tool was designed to identify priority pesticides at a very coarse, national level. The overall aim of this current project requires a more refined catchment- or farm-level approach and is focussed on aquatic ecological health as opposed to human health.

The aim of the risk-based indicator developed in this deliverable is to:

- 1. Assist in reducing the impacts of agricultural chemicals in water resources by allowing farmers and catchment managers to easily assess the risk of different pesticides to the environment and therefore choose pesticides that pose the least risk to the environment.
- 2. Assist catchment managers in prioritising high-risk pesticides within a catchment based on the crops produced in the catchment.

A key component of this deliverable was to develop a risk assessment methodology that compliments existing guidelines of pesticide application for different pests on different crops, as provided by Association of Veterinary and Crop Associations of South Africa (AVCASA) and CropLife South Africa. These pesticide application guidelines give recommended application rates for all pesticides (including insecticides, fungicides and herbicides) registered for use on a specific pest on a specific crop. For example, up to eight different pesticide are registered for use for control of false codling moth on peaches. These pesticides vary significantly in terms of their recommended application rate (Table 2-1), physicochemical properties and toxicity to aquatic biota (Table 2-2). Application rates are a key input to assessing the risk of a chemical to the environment. Using these application rates in combination with other parameters that influence exposure (i.e. physicochemical properties) and by using empirical pesticide transport equations, relative predictions of pesticide loading associated with use of the different pesticides will be made. Using a deterministic risk assessment approach, the relative risk of the different pesticides registered for use on a pest for a particular crop can be identified (i.e. identify which of the 8 insecticides registered for use on false codling moth pose the least risk to aquatic environment).

The output of this risk assessment is therefore an additional source of information to the guidelines provided by CropLife and AVCASA so that farmers can not only identify what pesticides are registered for use for a particular pest on a particular crop, but can then also identify which of these pesticides poses the lowest risk to the environment. Similarly, water resource/catchment managers could use the outputs to identify which high risk pesticides could be potentially used in a catchment based on the type of crops produced in the catchment.

Table 2-1: An example of pesticide application guidelines for controlling false codling moth on peaches.

Active Ingredient	Half- Life	KOC	Algae - Acute EC50 $(mg.L^{-1})$	Daphnia - Acute EC50 (mg.L ⁻	Fish - Acute LC50 $(mg.L^{-1})$
	(days)				
alpha-	35	57889	0.1	0.0003	0.0028
cypermethrin					
azinphos-methyl	10	1000	7.15	0.0011	0.02
beta-cypermethrin	10		56.2	0.00026	0.03
cypermethrin	60	76344	0.1	0.00015	0.00069
methomyl		25.2	60	0.0076	0.63
spinetoram	No Data				
triflumuron	22	2757	0.025	0.0016	320
zeta-cypermethrin	21	121786		0.00014	0.00069

Table 2-2: Physicochemical properties and toxicity endpoints for insecticides registered for control of false codling moth on peaches

2.4 GENERAL APPROACH

The approach adopted in the development of the indicator involves the development of an Excel based software tool that provides an indication of the relative risk of pesticides registered for the treatment of a specific pest on a specific crop. The tool enables the user to select a combination of pest and crop, for which a list of registered active ingredients is automatically generated. Based on the input of several easily obtainable input parameters, the tool provides a relative assessment of the risk of each pesticide to aquatic ecosystem health. This risk is estimated by calculating a PEC for each pesticide, which is compared to a toxicity concentration for each pesticide. The ratio of the PEC to the toxicity value, gives an indication of the risk of the pesticide and comparison of ratios for all pesticides provides a means to compare risks between pesticides.

The calculation of the PEC assumes contribution of pesticide loads derived from runoff and spray drift. The load of pesticide derived from each transport pathway is calculated by the spreadsheet, using simple, empirical formulae. These loads are expressed as a concentration through the input of basic stream dimension and flow parameters.

The spreadsheet relies on three databases, which have been developed as part of this deliverable and include the following:

- A database of active ingredients registered for use on specific pests for specific crops;
- A pesticide physicochemical database;
- A pesticide toxicity database;
- A database of crop-specific pesticide application rates.

Data used to develop each of the above-mentioned databases is discussed in more detail in sections below.

2.5 CALCULATION OF PEC

2.5.1 OECD runoff model

The OECD model as described in Section **Error! Reference source not found.** was used to estimate the percentage of the applied pesticide that was lost in runoff.

2.5.1.1 Data Requirements

Physicochemical data

Physicochemical properties greatly influence the fate and transport of pesticides in terrestrial and aquatic environments. These properties vary significantly from one pesticide to another, leading to differences in the mobility and exposure potential of pesticides in aquatic environments. The most important physicochemical properties required by the OECD model and that determine the fate and movement of pesticides in the soil, are the sorption coefficient of the active ingredient to organic carbon (*KOC*) and halflife (*DT50soil*) (Pionke and Chesters 1973; Gavrilescu 2005; Müller *et al.* 2007, Gassmann *et al.* 2015).

Environmental fate data should ideally be generated under conditions that would be expected to occur in the country/area where the pesticide is being registered. While this is often the case in countries from North America and the European Union, this is not necessarily so for other countries. For example, while this data is required for registration of pesticides in South Africa, the data is often generated in countries outside of South Africa. There is often a degree of reservation about using physicochemical data from more temperate climates as combinations of the chemical properties as well as site-specific environmental conditions (e.g. soil properties, temperature, etc.) also play a role in the fate and behaviour of pesticides (Daam and Van den Brink, 2010). These conditions vary greatly among different agro-ecological zones making the direct extrapolation of data between geographical regions very challenging (Ahmad and Kookana, 2007). However, Wauchope *et al.* (2002) found that while there is often variation in the Koc value of a specific pesticide (as an example), the values are adequate for discriminating between the relative mobility of a number of different pesticides. A study on the behaviour of three pesticides in South African soils reported similar *KOC* values to those reported in the international literature, while half-lives were generally longer in South African soils (Meinhardt, 2009). Other studies performed in South Africa have also shown good correspondence between Koc values and partitioning of pesticides between the sediment and water phase (Dabrowski *et al.*, 2002; Sereda and Meinhardt, 2005). These studies indicate that the international values provide a relatively good indication of pesticide behaviour in soils of South Africa and have been used successfully in predicting the relative mobility of pesticides under South African conditions (Dabrowski and Balderacchi, 2013).

Physicochemical data required for input into the OECD model was therefore obtained from the Pesticide Properties Database (PPDB, Lewis *et al.*, 2016). The PPDB is a comprehensive relational database of pesticide chemical identity, physicochemical, human health and ecotoxicological data. It has been developed by the Agriculture & Environment Research Unit (AERU) at the University of Hertfordshire for a variety of end users to support risk assessments and risk management. This data was captured in an Excel database from which the relevant data is retrieved for runoff calculations.

Pesticide Use Data

The quantity of pesticide applied is obviously an important factor in estimating losses into the environment. For the purposes of the indicator developed as part of this deliverable, two sources of pesticide use data were considered; estimated pesticide application rates based on market research and recommended pesticide application rates for products registered in South Africa.

Market Research Application Rates for Pesticides in South Africa

Pesticide use data for South Africa was purchased from GfK Kynetec (http://www.gfk.com/gfk-kynetec/), an international market research company. Data was provided as the aggregated total amount of active ingredient applied to all crops as well as the disaggregated application per crop. This data provides estimates on application rates (kg/ha) per active ingredient per crop. While this data is generated from market research, the US Geological Survey has sufficient confidence in the data provided by GfK Kynetec to use the data for estimating pesticide use in the USA as part of their National Water Quality Assessment Programme (Thelin and Stone, 2013).

Recommended Pesticide Application Rates for Pesticides Registered in South Africa

The risk indicator also allows for customised input of pesticide application data. This is particularly useful for farmers that have to deal with specific pests affecting their crop. Recommended rates of application for different pests on different crops can be obtained from guidebooks produced by AVCASA and CropLife. These books are published annually and list all pesticides registered for use on specific pests per crop. Application rates for each of the products can be estimated from the guidebooks according to specific formulae for insecticides, fungicides and herbicides, respectively.

The risk of pesticide products to aquatic biota are calculated based on the exposure to the active ingredient of the product. Pesticide products contain a certain concentration of the active ingredient, which is normally expressed as g.L-1, or g.kg-1 of the pesticide product. For example, the product Dursban contains the active ingredient chlorpyrifos at a concentration of 480 g.L⁻¹. In order to derive a PEC for the active ingredient it is important to calculate the application rate of the active ingredient and not of the product.

For example, the product Goal that contains oxyfluorfen as the active ingredient is registered for use as a herbicide on onions. The recommended dosage is three L.ha-1. The concentration of oxyfluorfen in Goal is however 240 g.L⁻¹. The application rate of the active ingredient is therefore calculated as follows:

$$
3L.ha^{-1} \times 240 \text{ g.}L^{-1} = 720 \text{ g.}ha^{-1} = 0.72 \text{ kg.}ha^{-1}
$$

Calculations for each of the active ingredient classes (fungicides, herbicides and insecticides) are explained in more detail in the sections below.

Herbicides

The recommended dosage for herbicides is provided as a unit of the volume or weight of the pesticide active ingredient applied per hectare (i.e. L.ha-1 or kg.ha-1). The recommended application rate of the active ingredient is therefore calculated as follows:

$$
RAR = \frac{GAI}{1000} \times RD
$$

Where *RAR* is the recommended application rate of the active ingredient (kg.ha⁻¹); *GAI* is the gram active ingredient per litre (g.L-1) or per kg (g.kg-1) of the applied product, and *RD* is the recommended dosage of the product $(L.ha^{-1}$ or kg.ha⁻¹).

Fungicides & Insecticides

Recommended dosage for fungicides and insecticides is typically provided as a unit of volume (ml) or weight (g) of the product mixed per 100 ℓ of water used to make up the total spray volume. The total spray volume per hectare is dependent on the method of application which can be a high volume (1000 L.ha-1 or higher), low volume (200 – 250 L.ha⁻¹) or aerial application (30 L.ha⁻¹). The recommended application rate of the active ingredient is therefore calculated as follows:

$$
AR = \frac{GAI}{1000} \times \left(RD \times \frac{SV}{100}\right)
$$

Where *RAR* is the recommended application rate of the active ingredient (kg.ha-1); *GAI* is the gram active ingredient per litre (g.L⁻¹) or per kg (g.kg⁻¹) of the applied product; *RD* is the recommended dosage per 100 ℓ of water (ml or g) and *SV* is the total volume of pesticide mixture applied per hectare (ℓ .ha⁻¹)

Geographical Data

Geographical inputs required for the model are easily obtainable from either field-based observations or desktop mapping resources (Table 2-3).

Data	Source	Reference	
		Directorate: National Geo-Spatial Chief Information	
Slope	Topographical Maps	(www.ngi.gov.za)	
Soil Type	WR90 Soils Map	Water Resources of South Africa, 2012 (WR2012)	
		http://waterresourceswr2012.co.za/resource-centre/	
Soil OC%	Agricultural Research	Department of Agriculture Forestry and Fisheries	
	Council	(http://daffarcgis.nda.agric.za/portal/home/)	
Vegetative Buffer	Field measurements		
Width	Google Earth Imagery		
Plant Interception	See Appendix 1	Linders <i>et al.</i> (2000)	

Table 2-3: Source of GIS data used to extract geographical information required to predict

The OECD runoff equation determines the total amount of runoff (mm) following a rainfall event based on whether the soil type is a sand or loam soil. The WR90 soil coverage can be used to identify soils according to these two categories based on the aggregation of classes as listed in Table 2-4.

2.5.2 Spray drift model

In Europe drift loadings are calculated based on drift values derived by the German BBA (Ganzelmeier *et al.*, 1995; Rautmann *et al.*, 1999), which divided crops into five groups (arable crops, fruit crops (orchards), grapevines, hops and vegetables/ornamentals/small fruit), with additional distinctions made between the early and late growth stages for fruit crops and grapevines and a crop height distinction for vegetables/ornamentals/small fruit. For each crop type and growth stage combination, experimental spray drift deposition data have been compiled as a function of distance from the edge of the treated field (Table 2-5). The 90th percentile drift values were calculated for each distance and used to generate a 90th percentile regression curve for each crop and growth stage combination. The figures in Table 2-5 are the percentage of an applied pesticide that is likely to move off target as spray drift. The application rate of the pesticide can be determined based on resources or methods described in 2.5.1.1.

Table 2-5: 90th percentile drift values based on crop type and distance from the point of application.

Drift values generated by the German BBA have been successfully validated in a number of studies conducted in the Lourens River catchment, South Africa. Deposition of azinphos-methyl and endosulfan was accurately predicted by the Ganzelmeier drift values at varying distances from the point of application (0, 5, 10 and 15 m) (Schulz *et al.*, 2001). In another study, drift deposition of azinphos-methyl in a stream adjacent to a pear orchard was accurately predicted by the Ganzelmeier drift values (Dabrowski *et al.*, 2005).

2.5.3 Toxicity data

Toxicity data required for assessing the relative risk of pesticides was obtained from the Pesticide Properties Database (PPDB, Lewis *et al.*, 2016). The PPDB is a comprehensive relational database of pesticide chemical identity, physicochemical, human health and ecotoxicological data. It has been developed by the Agriculture & Environment Research Unit (AERU) at the University of Hertfordshire for a variety of end users to support risk assessments and risk management. This data was captured in an Excel database from which the relevant data is retrieved for risk calculations.

2.5.4 Deterministic risk assessment

Deterministic methods are commonly used to assess risk the risk of pesticides to the aquatic environment. In this approach, a risk quotient (RQ) is calculated by dividing a point estimate of exposure (i.e. a single PEC value) by a point estimate of effects (i.e. relevant toxicity value). The calculation therefore integrates ecological effects (obtained during the exposure assessment) and exposure (pesticide use and fate and transport data) in quantifying risk. This ratio is a simple, screening-level estimate that identifies "risk" or "no risk" situations:

> *PEC/(Toxicity Value)<1 → No Risk PEC/(Toxicity Value)>1 → Risk*

The output is therefore a single point estimate of risk which could result in a simple "Yes" or "No" decision. A major disadvantage of this method is that a single exposure and effect endpoint is used to make a decision on the potential risk that could be expected to occur in a natural field situation. This incorporates a large amount of uncertainty into the risk assessment calculation, as there is inherently a large amount of variability in factors that influence both of these endpoints that may not be adequately considered in evaluating true risk. Single point estimates of exposure (i.e. PEC value) derived from environmental fate and transport models used in exposure assessment are particularly uncertain due to a number of reasons (Dubus *et al.*, 2003b):

Spatial and temporal variability of environmental variables (e.g. physicochemical properties, soil properties and climatic and geographical factors) that influence model results;

- x Uncertainty originating from difference in field sampling methods used to determine physical or chemical properties of pesticides;
- Uncertainty in spatially referenced data; and
- The choice of model used to predict environmental concentrations, with some studies indicating that the variability in model results due to model selection could be more significant than that due to input parameter variation.

From an effect perspective, different species exhibit differing sensitivity to chemical stressors. There is also intra-species variation depending on the life-stage of the test organism.

In summary, deterministic methods are relatively simple to execute and interpret and can be used to determine what is safe and is most likely protective of the environment. There is however a large amount of uncertainty associated with the method, it is not predictive and could also be too conservative (or overprotective) which could lead to certain beneficial products not being approved for use in agriculture.

2.6 METHODS

2.6.1 Estimating runoff loads

The percentage of applied pesticide lost as runoff (*L%runoff*) is estimated using the OECD model. The quantity of pesticide lost as runoff is calculated as follows:

$$
Q_{runoff} = \left(\frac{L\%_{runoff}}{100}\right) \times \left(\frac{PAR}{10000}\right)
$$

Where Q_{runoff} is the quantity of pesticide lost as runoff (kg.m⁻²), $L\%_{\text{runoff}}$ is the percentage of applied pesticide lost as runoff (percentage) and *PAR* is the Pesticide Application Rate (kg.ha⁻¹).

2.6.2 Estimating spray drift loads

The percentage of applied pesticide lost as spray drift (*L%spray*) is estimated according to the distance between the crop and water body and the crop type using the Ganzelmeier *et al.* (1995) tables (Table 2-5). The quantity of pesticide lost as spray drift is calculated as follows:

$$
Q_{spray} = \left(\frac{L\%_{spray}}{100}\right) \times \left(\frac{PAR}{10000}\right)
$$

Where Q_{spray} is the quantity of pesticide lost as spray drift (kg.m⁻²); *L%_{spray}* is the percentage of applied pesticide lost as runoff (percentage) and *PAR* is the Pesticide Application Rate (kg.ha-1).

2.6.3 Estimating the predicted environmental concentration (PEC)

The Predicted Environmental Concentration (PEC) is calculated as follows:

$$
PEC = \left(\frac{Q_{runoff} + Q_{spray}}{D}\right) \times 1000
$$

Where *PEC* is in mg/L and *D* is the depth of the adjacent waterbody (m).

2.6.4 Estimating the risk

The risk of the pesticide applied to a specific drop is calculated as follows:

$$
RQ = \frac{PEC}{TOX}
$$

Where *RQ* is the Risk Quotient (dimensionless) *PEC* is the Predicted Environmental Concentration (mg.L-1) and *TOX* is the acute toxicity endpoint (mg.L-1). The *RQ* can be defined as the acute toxic effect ratio for active ingredients in a watercourse adjacent to fields that are treated according to the estimated application rate. Risk is calculated for algae, invertebrates and fish. Risk categories are expressed according to Table 2-6.

Symbol	Risk Quotient	Description
	RQ > 1	High Risk
	1 > RQ > 0.1	Medium Risk
	0.1 > RQ	Low Risk

Table 2-6: Categories of risk of pesticide application to aquatic biota

2.7 RISK INDICATOR SOFTWARE

2.7.1 Data input tab

Two versions of the indicator have been developed. The Generic Risk Indicator can be used when the user has some knowledge of which crops are produced in a catchment but does not have any information of what specific pesticides are applied to the crop. The Specific Risk Indicator can be used when the user has some knowledge of which pesticides are applied to a particular crop in a catchment (including their application rate) or can also be used to assess the relative risk of several specific pesticides of interest.

2.7.1.1 Generic risk indicator

The Data Input Tab allows the user to add relevant information required for the calculation of the PEC and the associated RQ (Figure 2-1). These input parameters are easy to obtain using readily available resources (e.g. Google Earth, Cape Farm Mapper, etc.) or GIS data, and have been discussed in more detail in the previous sections. Based on the input of parameters in this tab, the PEC for all pesticides that are likely to be used on the crop is automatically calculated through reference to lookup pesticide use tables (as supplied by the GfK Kynetec database) included in the spreadsheet. The Generic Indicator therefore provides a list of pesticides (and associated risks) that are likely to be applied on a crop (as determined by market research).

Figure 2-1: Example of Data Input Tab showing the information that needs to be entered in order to estimate the relative risk of pesticides applied to apples on aquatic biota.

The definitions for all parameters are included in the spreadsheet (which can be seen by hovering the mouse cursor over the blue coloured cell) and are described as follows:

Crop Type: Distinguishes between vines, fruit trees, and field crops and early and late growth stages, which can be selected from a drop-down list.

Crop: The specific crop type for which pesticide use data is available can be selected from a drop-down list.

Total Area of Fields in Catchment (TAC): The total area (ha) of the crop in the catchment.

Total Area of Fields Adjacent to Stream (TAS): Total area (ha) of the crop that lies within 50 m of a stream.

Total Stream Length (TSL): The total length of the stream network (m) that lies within 50 m of crop. **Average Stream Depth:** The average depth of the stream network (m).

Total Perimeter of Fields Adjacent to Stream (TP): The total perimeter of the crop (m) that lies within 50 m of the stream.

Average Slope of Field: The average slope (percentage) of all fields in the catchment area.

Organic Carbon Content of Soil: The organic carbon content (percentage) of the soil in which the crop is cultivated

Average Distance Between Field and Watercourse (D): The distance (m) between the edge of the field adjacent to the stream and the edge of the stream closest to the field

Vegetative Runoff Buffer (B): The width of buffer (m) that is covered by vegetation that can potentially mitigate pesticide transport through slowing surface flow and providing an organic surface for adsorption. This width is distinct from the distance between the field and watercourse. For example, the distance between the field and the watercourse may be 20 m, however only 5 m of this width may be vegetated and act as a buffer to runoff.

Rainfall: The amount of rainfall. This can be the average rainfall per rainfall event or expressed as the average rainfall per rainfall event above 6 mm.

Soil Type: Distinguishes between sandy soils and loamy soils that can be selected from a drop-down list.

The risk indicator can be applied at the scale of a single field or at a larger scale that could include many fields along a river reach or catchment area. A diagram illustrating how the different input parameters can be calculated for scenarios including multiple fields is included in the 'Input Data' tab (Figure 2-2).

Figure 2-2: Diagram illustrating methods for calculating the data input parameters required for the Generic and Specific Risk Indicators.

The majority of inputs can be derived using relatively simple GIS mapping techniques including the following:

- Delineation of sub-catchments within a larger catchment area using a Digital Elevation Model (DEM) for the catchment area (Figure 2-3);
- Identification of crop types per sub-catchment area based on existing
- \bullet Total area crop in the catchment (Figure 2-3);
- Total area and perimeter of crop within 50 m of a watercourse (Figure 2-4)
- \bullet Total length of river within 50 m of the crop (Figure 2-5)
- The average distance between the crop and the watercourse (Figure 2-5 and Table 2-7)
- \bullet The slope of fields under a specific crop (Figure 2-6)
- \bullet The average slope of all fields under a specific crop (Table 2-7)

GIS analysis would need to be repeated for each crop type and sub-catchment of interest.

Figure 2-3: Example of a GIS map showing wheat grown in a sub-catchment of the Kars River, from which the TAC can easily be calculated.

Figure 2-4: Example of a GIS map showing wheat fields that fall within 50 m (dotted red lines) of watercourses in a sub-catchment from which the TAS and TP can be calculated.

Figure 2-5: Example of a GIS map showing the length of watercourses within 5, 10, 15, 20, 30, 40 and 50 m of adjacent wheat fields from which the average distance between fields and watercourses can be calculated (see Table 2-7).

Table 2-7: Example of calculating the average distance of a wheat fields from watercourses based on output of GIS analysis displayed in Figure 2-5.

	σ and σ		
Distance (D) between field and watercourse (m)	Stream Length (SL) (m)	SL x D	Average Distance between field and watercourse (m)
5	11745	58725	
10	1964	19640	
15	1987	29805	
20	2324	46480	$= 558050/28624$
30	4575	137250	$= 19.5$ m
40	3530	141200	
50	2499	124950	
TOTAL	286241	558050	

1Total Stream Length (TSL) as defined in Figure 2-2

Figure 2-6: Example of a GIS map showing the area of different slope categories of wheat fields in a subcatchment from which the total area of each slope category can be calculated and used to estimate the average slope of fields under wheat (see Table 2-8).

Slope (%)	Area (ha)	Slope x Area	Average Slopes (%)
2.5	120	300	
7.5	212	1590	$= 8903/892$
12.5	561	7013	$= 10%$
TOTAL	892	8903	

Table 2-8: Example of calculating the average slope of wheat fields based on the output of GIS analysis displayed in

2.7.1.2 Specific risk indicator

As mentioned above the Specific Risk Indicator can be used when the user has some knowledge of which pesticides are applied to a particular crop in a catchment or can also be used to assess the relative risk of several specific pesticides of interest. In addition to the catchment data input that is required for the Generic Risk Indicator an additional input box is provided where specific pesticides can be selected (up to ten active ingredients can be selected) and the application rate can be included (Figure 2-7). An application rate calculator is included in the spreadsheet for calculating application rates as per Section 2.5.1.1.

Figure 2-7: Example of additional data input required in the Specific Risk Indicator.

2.7.2 Risk indicator outputs

2.7.2.1 Generic risk indicator

The Generic indicator provides the following outputs:

- Pesticide Use & Risk Summary (Figure 2-8): Table showing:
	- o A summary the different pesticides typically applied to the crop and ranked according to their total application (kg).
	- o The application rate used in risk calculations;
	- o The relative risk of pesticides applied to the crop to algae, invertebrates and fish.
	- o Risks are differentiated between runoff and spray drift.
- Toxicity and Physicochemical Data (Figure 2-9): Table showing the toxicity and physicochemical data used in risk predictions; and
- Ranked Risk Table (Figure 2-10): Tables showing the relative risks of each pesticide to algae, invertebrates and fish, ranked from highest to lowest.

Figure 2-8: Example of the Pesticide Use & Risk Summary for peaches.

Figure 2-9: Example of the Toxicity & Physicochemical output table, providing relevant data used to determine the PEC.

Figure 2-10: Example of the ranked risks of pesticides to invertebrates for pesticides applied to peaches/

The outputs of the Generic Pesticide Risk Indicator are based on crop-specific application rates obtained from the pesticide use database as described in Section 1.2 and provided in the Appendix to this report.

The utility of the outputs can be summarised as follows:

- Provides a best guess of which pesticides could be applied in a catchment area where no detailed pesticide application information is available;
- Provides an indication of the relative risks these pesticides may pose to algae, invertebrates and fish;
- The outputs differentiate risks associated with runoff and spray drift and therefore provide guidance on monitoring strategies as well as which management interventions are most likely needed to control the source of pesticides in the catchment (i.e. spray drift will require different management interventions in comparison to runoff).
- The outputs can identify which pesticides are likely to be used in high quantities and the risk associated with these pesticides and can therefore be used to identify target pesticides for further monitoring/screening/management; and
- The indicator can be applied in different sub-catchments or tributaries in order to identify specific hotspot areas within a larger catchment area.

2.7.2.2 Specific indicator

For the Specific Risk Indicator, relative risks of only the selected pesticides are provided. The Specific Risk indicator provides the following outputs:

- Table showing the PEC for pesticides that originate from runoff and spray drift (Figure 2-11).
- Relative risk tables illustrating the relative risk of each pesticide to algae, invertebrates and fish (for runoff and spray drift) (Figure 2-12).

PECs $(\mu g/L)$			
Pesticide	Runoff	Spray Drift	
chlorpyrifos	1.474	9.208	
mancozeb	11.448	2.877	
fenarimol	0.536	0.460	
cypermethrin	0.011	0.767	
trifloxystrobin	0.548	0.959	
abamectin	0.001	0.017	
bupirimate	0.418	0.360	
lambda-cyhalothrin	0.001	0.192	
deltamethrin	0.007	0.384	

Figure 2-11: Example of the Specific Risk Indicator output providing an indication of estimated pesticide concentrations associated with runoff and spray drift.

Figure 2-12: Example of output provided by the Specific Risk Indicator.

The outputs of the Specific Pesticide Risk Indicator are based on user defined pesticides and associated application rates. The utility of the outputs can be summarised as follows:

• Provides a higher confidence estimate of relative risks based on more detailed knowledge of pesticide application programmes in the catchment;

- Provides an indication of the relative risks these pesticides may pose to algae, invertebrates and fish;
- The outputs differentiate risks associated with runoff and spray drift and therefore provide guidance on monitoring strategies as well as which management interventions are most likely needed to control the source of pesticides in the catchment (i.e. spray drift will require different management interventions in comparison to runoff);
- The indicator can be applied in different sub-catchments or tributaries in order to identify specific hotspot areas within a larger catchment area; and
- The indicator can be used to assess how changes in certain crop management inputs (i.e. increased buffer widths, vegetated buffer zones, etc.) can potentially affect the risk associated with the use of a pesticide;
- The indicator can be used to compare specific pesticides of interest with a view to identifying lower risk pesticides (i.e. for application to crops) or higher risk pesticides that might need to be included in a monitoring programme.

2.8 SOFTWARE PRODUCTS

Table 2-9 provides a summary of electronic outputs produced as part of the development of the crop specific risk indicator pesticide risk maps for South Africa.

File Name	Description
	Excel software tool designed to assess the relative risks of pesticides
GenericRLxlsx	applied to different crop types in catchments without any detailed
	knowledge of pesticide application programmes.
	Excel software tool designed to assess the relative risks of pesticides
	applied to different crop types in catchments with some knowledge of
SpecificRI.xlsx	pesticides used and their application rates. The tool can also be used
	to compare the relative risks of pesticides of interest.

Table 2-9: Summary of software products.

2.9 APPLICATION OF RISK INDICATOR

The software is currently in an initial trial phase and will be adapted and improved over the course of the project. In its current form, the following applications are envisaged:

- 1. Prioritisation of crop specific high priority pesticides for monitoring and management at a catchment scale;
- 2. Analysis of the risk to aquatic biota of different pesticides registered for use on specific pests targeting specific crops. This allows for improved decision making at a farm management scale; and
- 3. Changes in risk to aquatic biota can be assessed based on the implementation of specific farm management interventions (e.g. implementation of a vegetative buffer strip, increasing buffer width between fields and the edge of a watercourse, etc.).

An example of the application of the software can be viewed in Volume 1 of this research report.

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APPENDIX 1

Plant interception factors for different stages of crop growth according to Linders *et al.* (2000).

National pesticide use data for South Africa for the year 2014 (Data Source: Gfk Kynetec: www.gfk.com/gfk-kynetec)

Summary of active ingredient use for agricultural crops in South Africa for the year 2014 (Data Source: GfK Kynetec: www.gfk.com/gfk-kynetec)

